



Effects of Off-Coast Bluefin Tuna Fattening on Water Quality in the Eastern Mediterranean Sea

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Abstract

Bluefin tuna farming in the Mediterranean has been expanding since the mid-1990s. Expansion of tuna farming has been accompanied by widespread concerns about its impacts on the marine environment. The aim of the present study was to investigate impacts of a capture-based tuna farm located in the Gerence Bay on the water column and sediment.

With this aim, temperature, salinity, pH, dissolved oxygen, chlorophyll *a* and nutrients in the water column and organic carbon variables in the sediment were measured on a seasonal basis from 2011 to 2012. Samplings were made at the two reference stations and five stations around the cages. It was found that nutrient concentrations in the water column at the cage stations increased more than those of the reference stations but these differences were not statistically significant. Similarly, physico-chemical parameters and chlorophyll *a* in water column and organic carbon in sediment did not show detectable impact of tuna fattening. This was probably caused by strong currents present in the area, location of the cages away from the coast, hence high water depth and controllable feeding. It is believed that conducting monitoring studies is necessary to maintain sustainability of tuna fattening and determine the related potential impacts.

Keywords: Aquaculture, bluefin tuna, organic carbon, Mediterranean Sea, nutrients, off-coast farming.

Doğu Akdeniz’de Kıyı Ötesi Mavi Yüzgeçli Orkinos Besiciliğinin Su Kalitesine Etkileri

Özet

Akdeniz’de mavi yüzgeçli orkinos yetiştiriciliği 1990’ların ortalarından başlayarak gelişmiştir. Orkinos yetiştiriciliğinin yaygınlaşmasını izleyen süreçte, faaliyetin deniz ortamına etkileri ile ilgili önemli kaygılar oluşmuştur. Bu çalışmanın amacı, Gerence Körfezi’nde avcılığa bağlı kurulmuş olan bir orkinos çiftliğinin su kolonu ve sedimente olan etkilerinin araştırılmasıdır. Bu amaç doğrultusunda, su kolonunda sıcaklık, tuzluluk, pH, çözülmüş oksijen, klorofil *a* ve besleyici tuzlar ile sedimentte organik karbon değişkenleri 2011 ile 2012 yılları arasında mevsimsel olarak ölçülmüştür. Örneklemeler 2 referans istasyonu ve kafeslerin etrafındaki 5 istasyonda gerçekleştirilmiştir. Kafes istasyonlarında su kolonundaki nütrient konsantrasyonlarının referans istasyonlarından daha fazla artış gösterdiği fakat bu artışların istatistiksel olarak önemli olmadığı saptanmıştır. Benzer şekilde, su kolonundaki fiziko-kimyasal değişkenler ve klorofil *a* ile sedimentte organik karbon değerleri de orkinos besiciliğinin etkilerinin önemli seviyelerde olmadığını göstermiştir. Bu durumun çalışma bölgesinde mevcut olan güçlü akıntılar, kafeslerin kıyıdan uzak yerleşimi dolayısıyla derinliklerin fazla olması ve beslemenin kontrollü yapılmasından kaynaklandığı söylenebilir. Orkinos besiciliğinin sürdürülebilirliğini sağlamak ve potansiyel etkilerini belirleyebilmek için izleme çalışmalarının devam ettirilmesinin gerekli olduğu düşünülmektedir.

Anahtar Kelimeler: Akuakültür, orkinos, organik karbon, Akdeniz, nütrientler, kıyı ötesi yetiştiricilik.

Introduction

Rearing Atlantic bluefin tuna (*Thunnus thynnus* Linnaeus, 1758) in the Mediterranean has been expanding since the mid-1990’s. Bluefin tuna is one of the most demanding and expensive tuna species in the market. Its fishery is regulated by the International Commission for the Conservation of Atlantic Tunas

(ICCAT). The main growers of tuna in the Mediterranean are Italy, Malta, Spain, Croatia and Turkey. Other farms are located in Cyprus, Tunisia and Greece. Currently 54 firms are authorized to run bluefin tuna farming around the Mediterranean Sea. Tuna farming in Turkey has been started in 2002 and developed rapidly since then (Aksu, Kaymakçı-Başaran and Egemen, 2010). Although total output

capacity of the present 6 farms is assumed to be 6140 tons a year in Turkey today while nominal annual catch was 535 tons in 2012 (ICCAT, 2014).

Thunnus thynnus is considered a capture-based aquacultural species as farming activity entirely depends on stocking of wild caught fish. Wild tunas are caught by purse-seines, swum into towing cages and dragged slowly, usually at 1-1.5 knots where they are fed and kept for 4-8 months, which is called “fattening” (Ottolenghi, 2008). During fattening period tuna are fed mainly with small pelagic species including sardines, herring, mackerel and squid (Basaran and Özden, 2004). Food conversion ratios (FCR) are very high ranging from 15 to 30 (Aguado-Giménez and García-García, 2005; Vita and Marin, 2007).

Expansion of tuna farming in the Mediterranean has been accompanied by widespread concerns about impacts of the industry on the environment (Moraitis et al. 2013), the most important of which is unconsumed feeds and metabolic wastes due to fattening activity (Vezzulli et al. 2008). Fishing effort on wild tuna juveniles and pelagic fish stocks used for feed supply are other environmental issues (ICES, 2002). While studies on environmental effects of tuna farming are quite limited (Vita et al. 2004; Aguado-Gimenez et al. 2006; Vita and Marin, 2007; Vezzulli et al. 2008; Aksu, Kaymakçı-Başaran and Egemen, 2010), they gradually tend to increase (Vizzini and Mazzola, 2012; Moraitis et al. 2013; Mangion et al. 2014).

Impacts of tuna rearing on marine environment are expected to be greater than those of other farming activities such as processes for sea bass and sea bream in terms of high biomass rear and feed conversion ratios, nevertheless almost all the studies show that they are negligible and limited to the cage area (Vita and Marin, 2007; Vezzulli et al. 2008; Aksu, Kaymakçı-Başaran and Egemen, 2010). In addition, environmental effects of fish farming at floating cages varies with farm management, feed type, stocking density and the hydrography of the site (Wu, 1995), another reason for which is that the capture-based aquaculture of tuna is not a continuous process and lasts only four to eight months. After fattening has been processed, the sea may regenerate itself.

Mediterranean mariculture has been moved from small land-based operations to large farms scattered along the coastline and more recently to off shore in order to cope with outcomes of increased production (Grigorakis, 2011). Offshore farming is thought to be advantageous for both the fish farms and the environment considering higher water quality based on exposure of offshore locations to currents, waves and winds, which is therefore regarded as a means to overcome the problems caused by coastal fish farming (Vezzulli et al. 2008). Covering large marine areas, tuna farms are generally constructed off-coasts. Although there is no consensus on the term “offshore”, Holmer (2010) defines terms of ‘coastal’,

‘off coast’ and ‘offshore’ farming considering distance from the shore, depth, exposure to offshore impacts and wave height. Off-coast farming is known to take place in depths between 10 and 50m, especially at sheltered areas which are visible coastwise. Accordingly, the studied farm could be called an ‘off-coast’ farm.

Considering continues expansion of new off-coast farms around the Mediterranean it is necessary to recognize environmental impacts of this new sector so as to facilitate sustainable development of tuna fattening. However, studies dealing with the impacts of capture-based tuna on water quality operated at off-coast conditions are lacking at the present time. The aim of this study was therefore to observe likely effects of a capture-based tuna farm on water quality operated at an off-coast location in the Eastern Mediterranean.

Materials and Methods

Study Area

The study was conducted in Gerence Bay, SE Aegean Sea with an area of approximately 42000 m² and connected to the surrounding sea through a strait (Figure 1a). When the study was performed, the farm had two 66 m and two 50 m cages where 460 ton tuna was stocked. In order that tuna should be fattened, fishes were used such as sardine, herring, anchovy, mackerel and menhaden-type species. Feeding was performed under the inspection of scuba divers to finally decrease potential negative impacts of feed wastes.

Sampling and Analytical Methods

In order to find out impacts of tuna farming, water and sediment samplings were seasonally made at the two reference stations, the one being 200 m (R200) and the other 1km (R1000) west to 5 stations (cage stations; C1, C2, C3, C4 and C5) and related cages over the area (Figure 1b) on 1 July 2011, 4 October 2011, 23 March 2012 and 30 May 2012. Coordinates and depths of the stations are presented in table 1.

During the study, surface, mid and bottom water samplings were taken seasonally with a Nansen bottle from each station. Temperature and salinity were measured by SCT meter (YSI 30), pH by pH meter (YSI pH100), dissolved oxygen by oxygen meter (YSI 55) and transparency by Secchi disc. The water samples were taken to the lab via cold-chain. Variations of nitrite-nitrogen, nitrate-nitrogen, ammonium-nitrogen and phosphate-phosphorus in the water column were colorometrically measured using spectrophotometer (Parsons et al. 1984).

Seawater was filtered through GF/C filter paper to determine chlorophyll *a* and left to acetone extraction for 20 hours and measured absorbance of

the samples using spectrophotometer (Strickland and Parsons, 1972; Parsons *et al.* 1984).

Sediment samplings were taken by Van-veen grab at each station. Organic carbon level in every sediment sample was measured using dichromate oxidation method according to modified Walkley-Black titration (Gaudette *et al.* 1974).

Data Analyses

Kolmogorov-Smirnov test and Levene statistical

analyses were conducted for normal distribution convenience test of parameters and homogeneity test of variances, respectively. Significance test of the groups whose variances were found homogeneous was performed by Tukey test following ANOVA and Mann Whithney-U test following Kruskal-Wallis performed for groups whose variances were not homogenous to find out variations between the stations and between the seasons (Ergün, 1995; Sömbüloğlu and Sömbüloğlu, 2000). In addition, Principal component analysis (PCA) was applied to

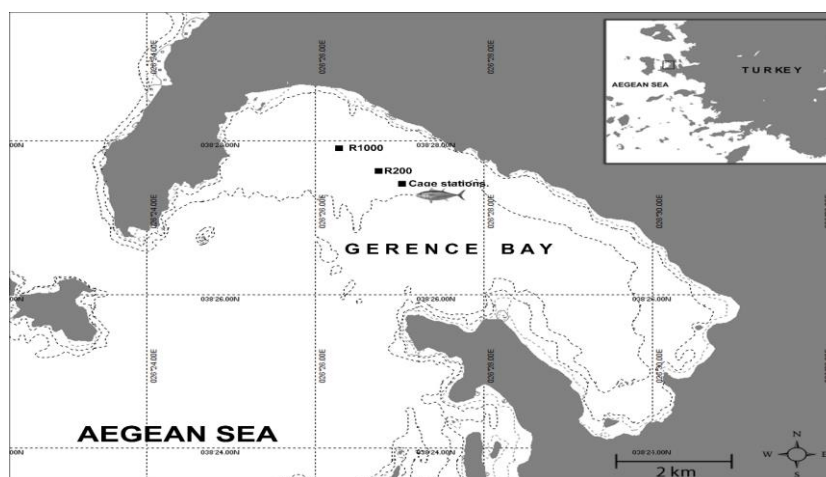


Figure 1a. Sampling stations in the study area.

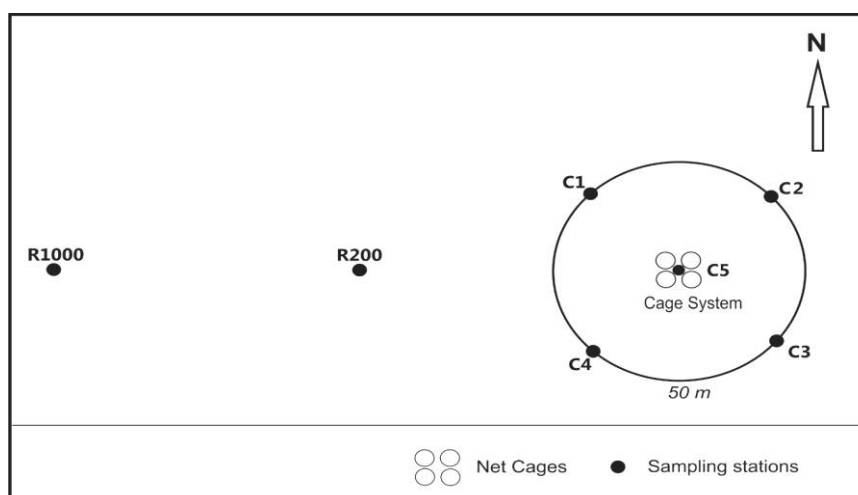


Figure 1b. Placement of the sampling stations (not drawn to scale).

Table 1. Coordinates and depths of the studied stations

Stations	Coordinates	Depths (m)
C1	38°27'20.3" N; 26°27'12.7" E	53
C2	38°27'22.3" N; 26°27'22.3" E	50
C3	38°27'11.9" N; 26°27'33.4" E	58
C4	38°27'10.8" N; 26°27'13.2" E	57
C5	38°27'18.0" N; 26°27'18.5" E	52
R200	38°27'25.1" N; 26°27'4.70" E	51
R1000	38°27'34.5" N; 26°27'45.9" E	48

evaluate to the temporal and spatial characteristics of water quality (Simeonov *et al.* 2003). PCA allows associations between variables, reducing the dimension of the data matrix. PCA was performed on normalized data using the factor procedure. All statistical analyses were done using the SPSS 16.0 software.

Results

Temperature

The samplings made in the Gerence Bay determined the lowest temperature as 14.7 °C in March 2012, at R1000 in the mid depth and the highest temperature as 23.9 °C in July on the surface water at C2 (Table 2). Means and standard error values for temperature were calculated as 19.61±0.33 °C in the Gerence Bay all the year round.

Salinity

The seasonal samples there showed the lowest salinity to be 35.0 ppt on surface water at C2 and C3 and mid water at C5 in May 2012 and the highest salinity 37.8 ppt in July 2011 at C3 in the mid water. Means and standard error values for salinity were found to be 36.34±0.07 ppt in the Gerence Bay all the year round.

pH

The lowest and highest pH's were measured as 7.68 and 8.28 in July 2011 at C4 on the surface and in March 2012 at R1000 in the mid waters, respectively. Means and standard error values for pH were found to be 8.06±0.02 in the Gerence Bay in 2011-2012.

Secchi Disc

The lowest and highest Secchi values were determined to be 6.05 m and 16.15 m at C3 and C2

stations in May 2012 and July 2011, respectively.

Dissolved Oxygen

The seasonal sampling in the area where the tuna farm is located found the lowest and highest dissolved oxygen concentrations to be 6.40 mg L⁻¹ and 8.30 mg L⁻¹ in the mid water at C3 and R1000 in October 2011 and March 2012, respectively (Table 2). Means and standard error values of dissolved oxygen were measured as 7.28±0.06 for the year concerned.

Nutrients

Nitrite-Nitrogen

The seasonal sampling found the lowest and highest nitrite-nitrogen to be 0.05 µM and 1.41 µM in the mid and bottom waters at C1 in July 2011 and May 2012, respectively. Table 3 presents seasonal minimum, maximum, average and standard values of nitrite-nitrogen measured at all the stations in Gerence Bay. Figure 2 illustrates annual means and standard errors.

Nitrate-Nitrogen

The lowest and highest nitrate-nitrogen concentrations were found to be 0.01 µM and 1.04 µM on surface and bottom waters at C1 and R200 in October 2011 and March 2012, respectively.

Table 3 presents seasonal minimum, maximum, average and standard values of nitrate-nitrogen measured at all the stations in Gerence Bay with Figure 2 illustrating annual means and standard errors.

Ammonium-Nitrogen

The minimum and maximum seasonal measurements of ammonium-nitrogen concentrations were found to be 0.06 µM and 2.78 µM on surface

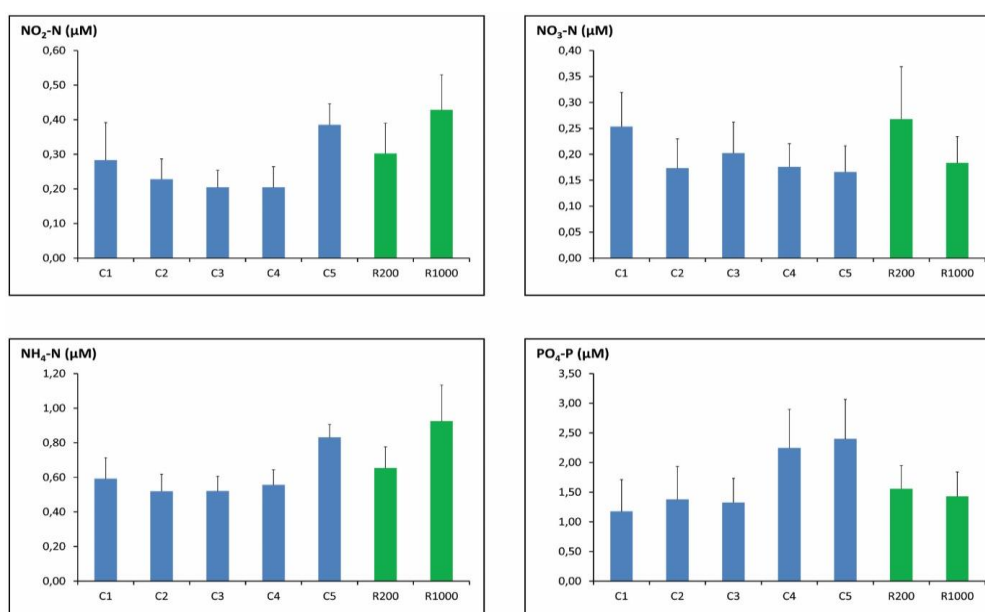
Table 2. Minimum, maximum, average and standard error values of physico-chemical parameters

Stations		Temperature (°C)	pH	DO (mg L ⁻¹)	Salinity (ppt)
C1	Range	14.8-23.3	7.79-8.21	6.65-8.12	35.1-36.9
	Mean ± SE	19.56±0.92	8.01±0.05	7.31±0.13	36.31±0.21
C2	Range	14.9-23.9	7.74-8.22	6.48-8.06	35.0-37.0
	Mean ± SE	19.73±0.96	8.05±0.05	7.15±0.15	36.22±0.20
C3	Range	14.8-22.5	7.86-8.24	6.40-8.23	35.0-37.8
	Mean ± SE	19.58-0.91	8.07±0.04	7.26±0.19	36.36±0.22
C4	Range	14.9-23.5	7.68-8.25	6.44-8.18	35.6-37.0
	Mean ± SE	19.61±0.95	8.06±0.05	7.23±0.17	36.36±0.15
C5	Range	15.1-22.7	7.85-8.24	6.56-8.01	35.0-37.0
	Mean ± SE	19.57±0.89	8.08±0.04	7.39±0.15	36.4±0.19
Ref 200	Range	15.1-22.6	7.83-8.25	6.50-8.17	35.2-37.1
	Mean ± SE	19.39±0.86	8.06±0.05	7.25±0.15	36.34±0.18
Ref 1000	Range	14.7-22.7	7.8-8.28	6.48-8.30	35.5-37.2
	Mean ± SE	19.86±0.88	8.07±0.05	7.36±0.16	36.38±0.17
All stations	Range	14.7-23.9	7.68-8.28	6.40-8.30	35.0-37.8
	Mean ± SE	19.61±0.33	8.06±0.02	7.28±0.06	36.34±0.07

Table 3. Minimum, maximum, average and standard values of nutrients

Periods		NO ₂ -N (µM)	NO ₃ -N (µM)	NH ₄ -N (µM)	PO ₄ -P (µM)
July 2011	Range	0.05-0.42	nd-0.24	nd-2.78	nd-6.74
	Mean ± SE	0.20±0.03	0.03±0.01	1.03±0.13	2.55±0.44
October 2011	Range	nd-1.18	0.01-0.61	0.06-0.97	0.92-6.74
	Mean ± SE	0.24±0.06	0.28±0.03	0.37±0.05	2.68±0.31
March 2012	Range	nd-0.48	nd-1.04	0.45-1.39	0.31-1.84
	Mean ± SE	0.20±0.03	0.45±0.05	0.69±0.05	0.54±0.08
May 2012	Range	nd-1.41	0.01-0.22	0.21-0.97	nd-6,43
	Mean ± SE	0.53±0.08	0.05±0.00	0.54±0.05	0.80±0.41

nd: non detected

**Figure 2.** Variation of annual nutrients (mean and standard error values)

and mid waters at C1 and R1000 in October 2011 and July 2011, respectively. Table 3 presents seasonal minimum, maximum, average and standard values of ammonium-nitrogen measured at all the stations in Gerence Bay.

Phosphate-Phosphorus

The minimum seasonal measurements of phosphate-phosphorus concentrations were found to be 0.31 µM at most of the stations in various depths in March 2012. The maximum measurement of phosphate-phosphorus concentration was found to be 6.74 µM on the bottom water at C1 in July 2011 and in the mid water at C2 in October 2011. Table 3 shows seasonal minimum, maximum, average and standard values of phosphate-phosphorus measured at all the stations in Gerence Bay.

Statistical analyses did not present significant differences for all nutrient elements between cage and

reference stations ($P>0.05$), only with important seasonal variations ($P<0.05$).

Chlorophyll a

The lowest and highest chlorophyll *a* values were found to be 0.38 µM and 5.67 µM on bottom waters at C1 and C2 in May 2012 and October 2011, respectively. Significant statistical differences were not observed for chlorophyll *a* between cage and reference stations ($P>0.05$).

Organic Carbon in Sediment

Organic carbon values ranged from 0.38 to 1.84 % at cage stations and 0.19 to 1.0 % at reference stations in sediment samples (Figure 3). The lowest and highest values were determined as 0.19 % and 1.84 % at R1000 and C5 stations in October 2011, respectively. Significant statistical differences were not observed for organic carbon between cage and

reference stations ($P > 0.05$).

Principal Component Analysis

Three PC's were found to account for 77 % of the variance in the whole data set.

The first PC, accounting for 32 % of total variance was correlated with pH, DO and phosphate. Excessive phosphate caused by fattening activities in the first PC enables organic matter to increase (phytoplankton blooming, feeding processes etc.). Decreases of DO values can be observed due to the presence and degradation of the organic matter in the medium. The second PC, accounting for 28 % of total variance was correlated with nitrite, nitrate and ammonium. The nutrient type component could be interpreted as representing influences of tuna rearing (aquacultural effects). Chlorophyll *a* accounted for the greatest loading for PC3, which explains 17 % of the total variance (Table 4).

Discussion

Effects of tuna fattening on the marine

environment is expected to be greater than that of farming of other fishes such as sea-bass and sea-bream because of its bigger biomass and the high feed conversion ratio (Aguado-Gimenez and Garcia-Garcia. 2005; Vezzulli *et al.* 2008). Nevertheless, effects of marine aquaculture vary based on the farm management, feed type, stocking density, and hydrodynamic conditions in the area concerned (Wu, 1995; Vita and Marin. 2007).

Values of temperature in the water column varying based on air temperature were found to be similar to each other except in March. The lowest values were normally observed in the winter sampling. Salinity and pH values varied seasonally with the highest in July 2011 and the highest in March 2012, respectively. The study found Secchi disk depth to have changed between 6.05 and 16.15 m with the lowest and highest in spring months and July when the primary productions were maximum and minimum, respectively. Aksu (2009) reported in the study conducted in Izmir Bay that Secchi disc value ranged from 3.15 to 9.45 m at the cage stations in fish farms in closed coves around it.

Concentrations of DO in coastal waters undergo

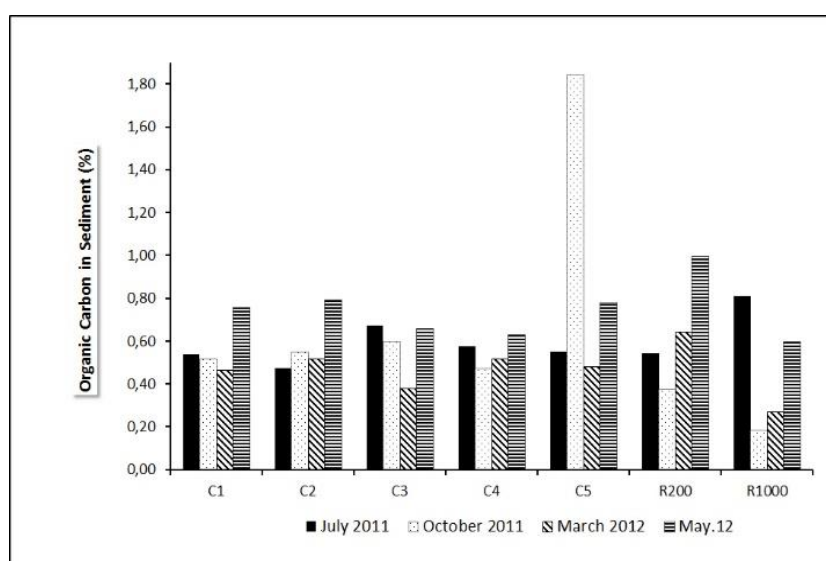


Figure 3. Variations of organic carbon values at the stations during the study.

Table 4. Loadings of experimental variables on the first three rotated principal components for the whole data set

Variable	PC1	PC2	PC3
DO	0,804	-0,332	-0,112
pH	0,735	0,525	0,362
Salinity	-0,256	-0,518	0,364
Chlorophyll <i>a</i>	0,543	0,246	-0,627
NO ₂	0,363	0,517	-0,531
NO ₃	0,490	0,766	0,012
NH ₄	0,162	-0,778	-0,410
PO ₄	-0,777	0,231	-0,468
Eigenvalue	2,55	2,23	1,34
Variance (%)	31,89	27,86	16,73

spatial and temporal change irrespective of aquacultural processes (Sangulin *et al.* 2010). Dissolved oxygen varied seasonally in the present study as well. The highest values were found at the reference stations during winter sampling whereas the lowest measurements were 6.4 mg L^{-1} at C4 station in October sampling, which is significantly higher than 5.0 mg L^{-1} , a critical value for aquatic organisms. Kaymakçı-Başaran *et al.* (2007) found DO values to range between 5.2 and 9.2 mg L^{-1} in the off-coast net cages at the fattening facility. Sangulin *et al.* (2010) reported that DO values typically varied between 7.03 and 7.62 mg L^{-1} at sea bass - sea bream and tuna farms with lowest value being 6.74 mg L^{-1} in autumn, which is consistent with those of the present study. The related authors (Aksu *et al.* 2010) in the long-term study at the same farm from 2005 through 2008 previously determined oxygen values to be $6.37 - 8.40 \text{ mg L}^{-1}$, which is also consistent with our present measurements ($6.40-8.30 \text{ mg L}^{-1}$).

Concentrations of nitrite, one of nitrogen forms did not show any significant differences between the stations but indicated that seasonal variations were comparatively effective and that concentrations varied from 0.05 to $1.41 \text{ } \mu\text{M}$. Nitrite-nitrogen values measured at the same farm changed between $\text{nd}-1.39$, $\text{nd}-2.51$ and $\text{nd}-1.08 \text{ } \mu\text{M}$ in 2005-2006, 2006-2007 and 2007-2008, respectively (Aksu *et al.* 2010). The highest value was found at the cage station while the study determined no significant differences between the stations as in our present study. Matijevic *et al.* (2006) determined lower nitrite values ($0.0-0.25 \text{ mmol m}^{-3}$) than those of the present study, reporting that nitrite measurements showed no significant statistical differences between the stations. Sangulin *et al.* (2010), also found lower nitrite values ($0.0002-0.0068 \text{ mg L}^{-1}$).

Nitrate-nitrogen values at the stations ranged from 0.01 to $1.04 \text{ } \mu\text{M}$ with no significantly statistical differences between them. Previous studies observed higher nitrate measurements such as $0.04-2.01 \text{ } \mu\text{M}$ (Aksu *et al.* 2010). The studies performed at other tuna fattening farms showed lower nitrate values than those at the present study (Matijevic *et al.* 2006; Sangulin *et al.* 2010).

Ammonium concentrations ranged from 0.06 to $1.69 \text{ } \mu\text{M}$ with no significant statistical differences between the stations. The same farm previously showed higher ammonium measurements which varied between $\text{nd}-1.34 \text{ } \mu\text{M}$, $0.09-3.53 \text{ } \mu\text{M}$ and $\text{nd}-2.48 \text{ } \mu\text{M}$ in 2005-2006, 2006-2007 and 2007-2008 (Aksu *et al.* 2010), respectively. Matijevic *et al.* (2006) reported that ammonium values ranged from 0.0 to 2.41 mmol m^{-3} without significant variations between the cage and reference stations. Sangulin *et al.* (2010) did not either find any significant differences between the cage and reference stations.

Nitrite, nitrate and ammonium concentrations were therefore found to indicate more effective in seasonal differences than those between the stations.

Phosphate-phosphorus concentrations fluctuated between 0.31 and $6.74 \text{ } \mu\text{M}$ with higher annual mean phosphate values at the cage stations. In addition, there were no significant variations between the cage and reference stations and phosphate measurements were higher than those in the previous studies (Aksu *et al.* 2010). Matijevic *et al.* (2006) reported lower phosphate values ($0.028-1.42 \text{ mmol m}^{-3}$). Matijevic *et al.* (2006) and Aksu *et al.* (2010) likewise found no significant differences between the cage and reference stations.

Chlorophyll *a* is a measure of photosynthetic phytoplankton biomass, changing based on solar ray and nutrients. The study showed that chlorophyll *a* fluctuated from 0.38 to $5.67 \text{ } \mu\text{g L}^{-1}$. Considering annual chlorophyll *a* values, the highest mean was observed as $2.52 \text{ } \mu\text{g L}^{-1}$ at C2 in the process of the study, for which no significant statistical variations were found between the stations. Pitta *et al.* (1999) measured highest chlorophyll *a* concentration as $3.5 \text{ } \mu\text{g L}^{-1}$ at the cage station in Ithaki farm. There are no environmental quality standard values for chlorophyll *a* around the Mediterranean coasts while chlorophyll *a* is a standard value as $10 \text{ } \mu\text{g L}^{-1}$ for North European waters to avoid eutrophication (Pitta *et al.* 1999) which was higher than $5.67 \text{ } \mu\text{g L}^{-1}$ by the present study.

One of the most important impacts of aquaculture on marine environment is that feeds not consumed by fish and their faeces have been accumulated below the cages on the sediment whose organic carbon concentrations is one of the remarkable variables in determining its quality (Karakassis *et al.* 2000; Hydland *et al.* 2005). Organic carbon concentrations in the sediment in the present study ranged from 0.19 to 1.84% . Statistical tests found negligible differences between the seasons and between the stations. On the other hand, the cage stations showed higher values than the reference stations. Annual means of the cage and reference stations were 0.64% and 0.55% , respectively. Aksu *et al.* (2010) reported organic carbon values ranging from 0.25 to 0.97% at the same farm. Such low values can be based on the fact that depths are high and current systems strong in the area. Lampadariou *et al.* (2005) reported that total organic carbon in the sediment at 7 different farms varied between 0.3 and 17.0% . Sangulin *et al.* (2010) reported that carbon values ranged from 0.53 to 2.58% at the farm where they performed their research on tuna. In that study, the cage stations found higher measurements than the reference stations did, which was attributed to local water currents causing smaller particles (faecal pellets) to be distributed over a wider area and large waste (excess feed) to sink directly just below cages.

In conclusion, nutrients in the water column and organic carbon concentrations of the sediments at the cage stations increased more than those of the reference stations but did not lead to any eutrophication risks due to the fact that the cages were

away from the coast and in deep water column with strong currents and feeding was conducted under control.

It is of great use to periodically perform the follow-ups in order to maintain sustainability of tuna fish fattening process in off-coast net cages and determine the related potential negative impacts.

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